

REVIEW ARTICLE

Diet Controls Uranium Intake and Aggravates Health Hazards

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Abstract

Uranium (U) is ubiquitously abundant in the environment. Significant amounts of U are applied on agricultural soils with mineral phosphorus fertilizers; thus, entering the food chain. The daily intake of U with solid food varies only slightly. It is lowest with a mixed standard diet (1.3 µg/day U) and highest with a vegan diet (2.0 µg/day U). It is the U content of tap and mineral water which determines the total U intake (1.7 – 7.1 µg/day U). Next to U speciation and amount of U entering the human body, an oversupply with phosphorus, and a critical supply with calcium and iron may amplify negative health effects. Kidneys are the prime target of a high P intake and U toxicity. The additional daily phosphorus intake by food phosphates peaked to 1000 mg/day in the last 20 years. This may add on an average 1.2 µg/day U and worst case 11 µg/day U to the solid diet. Toxicological studies suggest that damages of kidneys can be expected when the U content is as low as 0.1-0.4 µg/g U. The study provides a comprehensive overview of potential health hazards caused by dietary uranium intake in relation to nutritional habits.

Keywords: Calcium (Ca), drinking water, food phosphates, iron (Fe), phosphorus (P), uranium (U)

1. Introduction

Uranium (U) and phosphorus (P) are two elements which occur together in the animate and inanimate world. Phosphate rock is the basic material for the processing of mineral P fertilizers. P is essential for all living organisms and mineral P fertilization being vital for food security on a global scale. U has been extracted from phosphate rock until the late 1990s before it became unprofitable. Nowadays, 7-22 g/ha*yr U enter the food chain by different mineral P fertilizers if applied at a rate that replaces the mean P off-take of 22 kg/ha*yr P¹. An intake of U by natural food products is inevitable, however, the U uptake by crop plants is limited in general. U is highly mobile in agricultural soils and discharged to water bodies by run-off and leaching. The provisional guideline value for U in drinking water is 30 µg/L², in the U.S.A. EPA³ set a limit value of 20 µg/L U. In Germany, for the preparation of infant food a limit value of 2 µg/L and an action value of 20 µg/L in drinking water has been adopted^{4,5}. The U content of bottled mineral water worldwide has been compiled by Knolle⁶. Schnug et al.⁷ calculated a mean U intake of 2.5 µg/day by solid food. With a vegan and a carnivore dietary style the corresponding values are 1 and 4 µg/day U.

U poses a radiological risk and is a heavy metal poison that concentrates in bone, principally targeting lung and kidneys⁸. The incorporation of U into bones poses in particular a radiological risk as the retention time is about 200 days^{8,9}. In comparison, kidneys and liver have a high blood circulation so that the chemical toxicity prevails.

The dangers arising from the biochemical toxicity of U are generally considered to outweigh the risks from its radioactivity¹⁰. At this point it is important to note that it is the aim of this contribution to outline potential health hazards caused by the dietary intake of U. The impact of U on miners and Gulf War veterans exposed to dust or depleted U deserve separate attention; one effect of U in these humans was, however, chromosome damage^{11,12}.

U has three common isotopes, U-238, U-235 and U-234. With a specific activity of about 14 MBq/kg U has been considered to be a low

cancer risk¹³. The weighted energy dose is called 'equivalent dose' to a tissue and given in Sv (Sievert). The equivalent dose takes the type of radiation (α , β , γ) into account. Fertilizer-derived U that shows up in drinking water would statistically increase the annually added effective dose by 7 µSv for infants and 3 µSv for adults¹⁴. Schmitz-Feuerhake and Bertrell¹⁴ concluded that the radiological health risk of U entering the food chain by fertilization is only minor.

The uranyl ion exhibited anomalous genotoxic effects at low concentrations causing genomic and genetic damage in cell cultures at concentrations where there are no significant alpha emissions^{15,16} and caused cancer in laboratory animals¹⁷. U caused anomalous inflammation in lung, kidney, brain and other living tissue in rats and produced neurological effects in mice¹⁸. Thiebault et al.¹⁹ showed that U caused DNA damage and apoptosis in a concentration-dependent manner in rat kidneys. The molecular and cellular mechanisms responsible for the genotoxicity of U concentrate on vertebrates such as fish²⁰.

Neither the radiological, nor the chemical risk assessment of U delivered a satisfactory explanation for the occurrence of cancer. Busby and Schnug¹³ suggested that the so-called photoelectric effect multiplies the radiation of U. This means that U, e.g. bound to the DNA, absorbs γ -radiation from the environment resulting in the emission of α - and additional β -radiation so that the total radioactivity multiplies. DNA saturated with U would absorb >55,000 times more of the surrounding γ -radiation than the DNA itself. The photoelectric effect has been described in radiation therapy of cancer employing gold nano-particles²¹. Further damage may result from the bystander and low dose effect (Bishop 2005, citing Busby and Schnug¹³). Radiated cells signal to neighboring cells, which did not receive direct radiation and thus enhance damages. A detailed description of these effects is provided by Busby and Schnug¹³.

Not only the dietary style, but also dietary habits changed significantly during the past decades. The result is an increased serum phosphate level, which was determined twice

as often in socially weaker strata where people consumed preferably processed and convenience food products. For patients with a mild to moderate kidney insufficiency the risk of hyperphosphatemia was increased up to 2.7-fold in the lowest income group²². Kidneys are a target organ of high P intake; regulatory mechanisms result in the demineralization of bones and excretion of Ca (nephrocalcinosis)²³. Several reviews highlight the impact of high P intake on health hazards^{24,25}. The reason for an increased P intake is the use of food phosphates in processed food products. Their use is regulated in Europe in EU regulation No 231/2012 for maximum quantities of phosphates added during food processing²⁶. However, it is not so much the higher U intake by processed food products as combination toxicological aspects of a concomitant over-proportional P intake that poses a yet neglected risk of health hazards⁹.

2. Potential health hazards of dietary uranium intake

U is taken up with solid food and water. In surface water the U-species U(V), UO_2^+ und U(VI) can be found, which bind to complexes with sulfate, carbonate and hydroxyl-ions and phosphates. In plants U can be found as U(VI) in phosphate complexes irrespective of the U speciation in soils²⁷. Thus U ingested with solid food is less soluble and will be less absorbed in the human body as for instance soluble compounds in drinking water. About 0.2% of hardly soluble U species are absorbed in the human body²⁸, while it is 2.0% of the diet²⁹. This stresses the significance of the U intake by water as an integral part of nutritional intervention.

The prime targets of U toxicity are kidneys and bones². Kidneys as the U concentration is highest during excretion and bones because they act as sinks for U. Typical, dose-dependent U-induced renal injury caused in animals are lesions in the outer stripes of the outer medulla^{30,31}. The proximal convoluted tubules seem to be affected in particular³². Foulkes³³ stressed that it is not possible to define a general critical U-content for kidneys

as reactions are compartment-specific; damages can be expected when the U content is as low as 0.1-0.4 $\mu\text{g/g}$ U. The same author stressed that a NOAEL (no-observed-adverse-effect-level) should not exceed 0.3 $\mu\text{g/g}$ U as the transferability of animal studies to humans is restricted. Vicente-Vicente et al.³⁴ classified a declining sensitivity against U toxicity in the order rabbit > rat > guinea pig > pig > mouse > dog > cat > human. Chandrajith et al.³⁵ associate the increasing number of chronic kidney disease (CKD) patients of uncertain aetiology with an increased exposure to U for which Schnug and Lindemann³⁶ (2006) surmise mineral waters as a potential source. Schnug et al.³⁷ showed that the number of patients suffering from chronic kidney failure from 1986-2004 in Austria, Germany and the U.S.A. increased by 5% and this rise coincided with an increasing consumption of mineral water, which multiplied the daily U intake. A daily U intake of 1 $\mu\text{g/g}$ U results in a U concentration of 12 $\mu\text{g/g}$ U in kidneys and saturation of DNA with U¹³. In a clinical study Banning and Benfer³⁸ found a weak, however, significant relationship between the U content in drinking water and the occurrence of tumors and liver diseases. The liver bears about 16% of the U loads in humans³⁹.

Currently, statistical data about the dose/effect relationship of dietary uranium intake and human health are not available as awareness of its significance has just started. Nevertheless, circumstantial evidence has been collected that dietary habits and style control the personal daily uranium intake and the associated health risks they pose. In addition, the daily supply with minerals, in particular P, is obviously causally linked to U toxicity in the human body.

2.1 Impact of the mineral nutritional status on U human toxicity

The significance of the P, calcium (Ca) and iron (Fe) status for potential U-related health effects have been paid little attention so far. The Ca level in the blood plasma influences U speciation. U can be found in the protein and

non-protein fraction of blood and is bound to erythrocytes⁴⁰. This process is pH-dependent and declines steeply from pH 7 with the strongest binding capacity to pH 6 as carboxyl groups and lipoproteins dissociate less intensely. Uranyl-bicarbonate, $[\text{UO}_2(\text{HCO}_3)_4]^{2-}$ is the most stable compound because of the low dissociation constant and is excreted from the plasma via kidneys. U is abundant in the inorganic part of the blood in a bicarbonate complex where it replaces Ca and/or Mg. In comparison, U can be found in the protein part of the blood as uranyl-albuminate. In the blood plasma the product of $\text{Ca} \cdot \text{P}$ is almost constant. Under conditions of hyperphosphatemia the Ca level is low and sensitivity against U high⁴¹.

In the pH range of 5-7 UO_2^{2+} binds quickly to α -amylase and reduces its activity⁴². The same authors stress that Ca is crucial to counteract this effect. It is also the Ca level in the plasma that influences U speciation⁴³. Interesting in this context is that U destroys intact α -helix structures of amylase, hemoglobin and blood serum albumin^{42,44,45}. Computer simulation suggested that in a concentration dependent manner UO_2^{2+} -complexes are formed. UO_2 binds for example to PO_4 , and promotes incorporation into bones. Thus it is fair to assume that a higher dietary P and U intake poses an additional health hazard compared to a low U intake and P intake that meets the physiological demand.

The cytotoxicity of U in liver cells was closely related to the extracellular P concentration⁴⁶. Liu et al.⁴⁷ showed that uranyl-nitrate is transformed into uranyl-phosphate in liver cells, which has been suggested to cause cytotoxicity³⁵. Inorganic phosphate seems to actively influence metabolic cell processes⁴⁸. An elevated phosphate level increased cell proliferation and Fra-1 and osteopontin genes were expressed stronger. *In vitro*, increasing phosphate levels were linked to tumor expressing genes and *in vivo* the same authors found in dependence on dietary P intake next to elevated P levels in the serum and osteopontin, a reduced Ca content and an about 50% higher number of skin papilloma.

U absorption increases proportionally with the amount of U ingested⁴⁹. The U concentration in human hair is a direct, linear function of the U content in drinking water and indicator for U absorption by the human body⁵⁰. U absorption may increase manifold under conditions of Fe deficiency, even, up to 50-fold if Fe^{3+} is supplemented⁵¹. Both elements are absorbed by the same transporter (DTM1) and regulatory processes seem to be related to the redox behavior of U and P in such way that both a high external Fe supply and Fe deficiency may yield an increased U absorption⁴⁹.

3. Daily intake of uranium in relation to dietary style and dietary habits

Hassoun⁵² developed a model in which the considerably variable food sources of human diets were assembled in affiliated categories in order to achieve a low bias from mineral intake by solid food. Her standardized mixed diet is based on an energy requirement of 2000 kcal/day⁵³ and follows the rules of the trophic pyramid. Hassoun⁴¹ diversified the healthy mixed diet in three additional dietary styles: ovo-lacto vegetarian, vegan and carnivore. A standard diet supplies the smallest amount of U with solid food to the human body (1.3 $\mu\text{g}/\text{day}$ U) whereas a vegan diet delivers the highest amount with 2.0 $\mu\text{g}/\text{day}$ U⁵⁴. In comparison to solid food, consumers of low and highly mineralized bottled water have the lowest (1.72 and 1.90 $\mu\text{g}/\text{day}$) and consumers of German and world bottled mineral waters the highest U intake with 6.16 and 7.08 $\mu\text{g}/\text{day}$, respectively⁵⁴. The contribution of drinking water to the daily U intake of a tap water drinking standard diet consumer is 71.8% with a range of 56.7% for highly mineralized bottled water to 84.3% for a world bottled mineralized water consumer. The maximum reduction potential for the daily U intake through a change of water drinking and dieting habits is -67%⁵⁴. Consumers ingest significant amounts of U with mineral and potable water while the intake by solid food is distinctly lower and shows a much smaller bandwidth⁵⁴. The recommended upper daily P intake varies between 700 and 1250 mg⁵⁵. In general, the P concentration of plant products is only half as

high as that of animal products⁵⁴. The contribution of liquids to the daily P intake is negligible with <0.01% so that only a change of the dietary style may help to reduce the P intake by one third⁵⁴.

The term dietary habits reflects the consumption of natural food products versus processed and convenience foods. For the preparation of the latter food phosphates are added in varying amounts; maximum quantitative limits are defined in EU regulation No 231/2012 for the European Union²⁶. For

example, non-alcoholic, aromatized drinks may contain up to 700 mg/L P₂O₅ in form of phosphoric acid (E338), cheese spread 20 mg/kg and processed meat products up to 5 mg/kg P₂O₅ in form of potassium-phosphate (E340), oven-ready flour up to 20 g/kg P₂O₅ as di-magnesium-phosphate (E343). In Table 1 the descriptive statistics of 39 food phosphate samples are summarized. Variations in the total P content reflect different phosphate forms.

Table 1. Descriptive statistics of U (µg/g) in food phosphates (n=39; data extracted from Windmann⁹).

Element	Minimum	Maximum	Mean	Standard deviation
P	19780	265135	137521	68671
U	<LLD	1.90	0.257	0.379
P:U	0.08	11.03	2.031	2.100
<i><LLD – Lower Limit of Detection</i>				

The highest U content was 1.90 µg/g and exceeded the suggested limit value of 1.00 µg/g U⁹. Based on the data provided in Table 1 an additional dietary P intake of 1000 mg P by processed food products may add worst case up to 11 µg U to the solid diet⁸. Using the geometric instead of the arithmetic mean values which are more robust against extremes the P and U contents in the same food phosphate samples were 117,183 µg/g P and 0.14 µg/g U, respectively. This equals a daily additional U intake of 1.2 µg U⁹. These simple calculations demonstrate that potentially an excessive P intake by processed and convenience food products can be alleviated if a low-meat diet is selected.

4. Conclusions

Changes in lifestyle caused changes in the daily intake of U and P almost unnoticed as the

annual consumption of mineral water and processed food has been continuously increasing during the last decades^{6,56}. Phosphates are essential for the production of food and human health. And phosphates are the most significant source for the contamination of the food chain with undesired elements, especially U. Negative health effects of an increased U intake seem to be systematically amplified by an increased P intake and a reduced P intake is an integral part in chemoprevention⁴⁸. Other mineral nutrients unfolding crosstalk with U absorption and human toxicity are Ca and Fe. Following the precautionary principle, individual health care should target on a balanced intake of minerals, especially with P, Ca and Fe and an efficient limitation of the U intake through a conscious selection of drinking water and food products.

5. References

1. Kratz, S., F. Knappe J, Rogasik J, Schnug E. 2008. Uranium balances in agroecosystems. In: De Kok LJ, Schnug E, editors. Loads and fate of fertilizer derived uranium. Leiden: Backhys; 2008. p. 179- 189.
2. WHO. 2012. Uranium in drinking water, Background document for development of WHO Guidelines for Drinking-water Quality, WHO/SDE/WSH/03.04/118/Rev/1.
3. EPA. National Primary Drinking Water Regulations. Radionuclides 2000; 40CFR Part 141 (65FR 76708, Dec 7th 2000).
4. MTVO. 2006. Mineral- und Tafelwasser-Verordnung vom 1. August 1984, BGBI. I S. 1036, zuletzt geändert durch die Vierte Verordnung zur Änderung der Mineral- und Tafelwasser-Verordnung v. 1. 2006 Dezember 2006, BGBI. Bonn 11. Dezember, 2006. Teil 1 Nr. 56, 2762- 763. [online] <http://www.gesetze-im-internet.de/min_tafelwv/> (last accessed on 9 January 2020).
5. TrinkwV. Erste Verordnung zur Änderung der Trinkwasserverordnung vom 3. Mai 2011, Bundesgesetzblatt Jahrgang 2011 Teil 1 Nr. 21. Bonn: 11. November 2011, 748-774.
6. Knolle F. A contribution to occurrence and origin of uranium in German mineral and tap waters. Braunschweig, Germany; PhD thesis, Technical University 2008.
7. Schnug E, Birke M, Costa N, Knolle F, Fleckenstein J, Panten K, Lilienthal H, Haneklaus S. Uranium in German mineral and tap waters. In: De Kok LJ, Schnug E, editors. Loads and fate of fertilizer derived uranium. Leiden: Backhys; 2008. p. 91-110.
8. Royal Society. The Health hazards of depleted uranium munitions. Part I, London: The Royal Society, 2001.
9. Windmann H. A contribution to the risk assessment of uranium contamination in the food chain through phosphate containing fertilizers, feed and food additives. Braunschweig, Germany; PhD thesis, Technical University 2020.
10. Milvy P, Cothorn CR. Scientific background for the development of regulations for radionuclides in drinking water. In: Cothorn CR, Rebers P, editors. Radon, Radium and Uranium in Drinking Water. Chelsea, Michigan: Lewis Publishers 1990. P.1-16.
11. Zaire R, Notter M, Thie, E. Unexpected rates of chromosome instabilities and alteration of hormone levels in Namibian uranium miners. *Radiation Research* 1997; 147:579-584.
12. Schroeder H, Heimers A, Frentzel Beyme R, Schott A, Hoffmann W. Chromosome aberration analysis in peripheral lymphocytes of Gulf War and Balkans War veterans. *Radiation Protection Dosimetry* 2003; 103: 211-219.
13. Busby C, Schnug E. Advanced biochemical and biophysical aspects of uranium contamination. In: De Kok LJ, Schnug E, editors. Loads and fate of fertilizer derived uranium. Leiden: Backhys; 2008. p. 11-22.
14. Schmitz-Feuerhake I, Bertell R 2008 Radiological aspects of uranium contamination. In: De Kok LJ, Schnug E, editors. Loads and fate of fertilizer derived uranium. Leiden: Backhys; 2008. p. 1-10.
15. Miller AC, Stewart M, Brooks K, Shi L, Page N. Depleted uranium catalyzed oxidative DNA damage: absence of significant alpha particle decay. *Journal Inorganic Biochemistry* 2002; 91: 246-252.
16. Smirnova VS, Gudkov SV, Shtarkman IN, Chernikov AV, Bruskov VI. The genotoxic action of uranyl ions on DNA in vitro caused by the generation of reactive oxygen species. *Biofizika Akademija Nauk SSSR* 2005; 50:456-463.
17. Miller AC, Mog S, McKinney L, Lei L, Allen J, Xu J, Page N. Neoplastic transformation of human osteoblast cells to the tumorigenic phenotype by heavy metal-tungsten alloy particles: induction of genotoxic effects. *Carcinogenesis* 2001; 22:115-125.
18. ENVIRHOM 2005. Bioaccumulation of

- radionuclides in situations of chronic exposure of ecosystems and members of the public. Progress Report No 2. Report DRPH 2005-07, Fontenay aux Roses, France: IRSN, 2005.
19. Thiebault C, Carrière M, Milgram S, Simon A, Avoscan L, Gouge B. Uranium Induces Apoptosis and Is Genotoxic to Normal Rat Kidney (NRK-52E) Proximal Cells. *Toxicological Sciences* 2007; 98(2):479–487.
 20. Henner P. 2008. Bioaccumulation of radionuclides and induced biological effects in situations of chronic exposure of ecosystems – an uranium case study. In: De Kok LJ, Schnug E, editors. Loads and fate of fertilizer derived uranium. Leiden: Backhys; 2008. p. 23-32.
 21. Meshabi A. A review on gold nanoparticles radiosensitization effect in radiation therapy of cancer. *Reports of Practical Oncology & Radiotherapy* 2010; 15:176-180.
 22. Gutiérrez OM, Isakova T, Enfield G, Wolf M. Impact of poverty on serum phosphate concentrations in the third national health and nutrition examination survey. *Journal of Renal Nutrition* 2011; 21:140-8. DOI: 10.1053/j.jrn.2010.03.001.
 23. Weiner ML, Salminen WF, Larson PR, Barter RA, Kranetz JL, Simon GS. Toxicological review of inorganic phosphates. *Food and Chemical Toxicology* 2001; 39:759–786.
 24. Cooke A. Dietary food-additive phosphate and human health outcomes. *Comprehensive Reviews in Food Science and Food Safety* 2017; 16:906-1021.
 25. Ritz E, Hahn K, Ketteler M, Kuhlmann MK, Mann J. Phosphate additives in food – a health risk. *Deutsches Ärzteblatt International* 2012; 109:49-55.
 26. EU. Commission Regulation (EU) No 231/2012 of 9 March 2012 laying down specifications for food additives listed in Annexes II and III to Regulation (EC) No 1333/2008 of the European Parliament and of the Council. Brussels: L83/1. 2012.
 27. Günther A, Bernhard G, Geipel G, Reich T, Roßberg A, Nitsche H.. Uranium speciation in plan. *Radiochimica Acta* 2003; 91:319-328.
 28. Spencer H, Osis D, Isabel M. Measured intake and excretion patterns of naturally occurring ²³⁴U, ²³⁸U, and calcium in humans. *Radiation Research* 1990; 124:90-5.
 29. Zavodska L, Kosorinova E, Skerbakova L, Lesny J. 2019. Environmental chemistry of uranium. HU ISSN 1418-7108: HEJ Manuscript no.: ENV-081221-A. [online] <http://heja.szif.hu/ENV/ENV-081221-A/env081221a.pdf>. Last accessed on 9 January 2020.
 30. Homma-Takeda S, Terada Y, Nakata A., Sahoo SK, Yoshida S, Ueno S, Inoue M, Iso H, Ishikawa T, Konishi T, Imaseki H, Shimada Y. Elemental imaging of kidneys of adult rats exposed to uranium acetate. Nuclear instruments & methods in physics research section b-beam interactions with materials and atoms 2009; 267:2167-2170. DOI: 10.1016/j.nimb.2009.03.082
 31. Leggett RW. The behaviour and chemical toxicity of U in the kidney: a reassessment. *Health Physics* 1989; 57(3):365–383.
 32. Domingo JL. Chemical toxicity of uranium. *Toxicology and Ecotoxicology News* 1995; 2(3):74–8.
 33. Foulkes EC. The Concept of Critical Levels of Toxic Heavy Metals in Target Tissues. *Toxicology* 1990; 20:327-339.
 34. Vicente-Vicente L, Quiros Y, Pérez-Barriocanal F, López-Novoa JM, López-Hernández FJ, Morales AI. Nephrotoxicity of Uranium: Pathophysical, Diagnostic and Therapeutic Perspectives. *Toxicological Sciences* 2010; 118:324-347.
 35. Chandrajith R, Seneviratna S, Wickramaarachchi K, Attanayake T, Aturaliya TNC, Dissanayake CB. Natural radionuclides and trace elements in rice field soils in relation to fertilizer application: study of a chronic kidney disease area in Sri Lanka. *Environmental Earth Sciences* 2010; 60:193-201.
 36. Schnug E, Lindemann I. Verringerung der Strahlenbelastung durch bewusstes Konsumverhalten bei Trinkwässern.

- Strahlentelex Elektrosmog Report* 2006; 476-477:4-5.
37. Schnug, E, Steckel H, Haneklaus S. Contribution of uranium in drinking waters to the daily uranium intake of humans - A case study from Northern Germany. *Landbauforschung Völkenrode* 2005; 55:227-236.
 38. Banning A, Benfer M. Drinking Water Uranium and Potential Health Effects in the German Federal State of Bavaria. *International Journal of Environmental Research and Public Health* 2017; 14: 927. DOI: 10.3390/ijerph14080927.
 39. U.S. Health Dept. 2013. Toxicological profile for uranium, U.S. Department of Health and Human Services, Agency for Toxic Substances and Diseases Registry.
 40. Chevvari S, Likhner D. Complex Formation of Natural Uranium in Blood. *Journal of Radiology and Medical Imaging* 1968; 13:53-7.
 41. Anonymous. Der Calcium- und Phosphathaushalt [cited 27 April 2021]. Available from: <https://www.lecturio.de/magazin/der-calcium-und-phosphathaushalt/>.
 42. Barkleit A, Hennig C, Ikeda-Ohno A. Interaction of uranium (VI) with α -amylase and its implication for enzyme activity. *Chemical Research in Toxicology* 2018; 31:1032–1041. DOI: 10.1021/acs.chemrestox.8b00106.
 43. Bing D, Heyi W, Shubin J, Junge M. The biodistribution of uranium in mice. *Environmental Chemistry* 2011; 30:1247-1252.
 44. Ali M, Kumar A, Kuma M, Pandey BN. The interaction of human serum albumin with selected lanthanide and actinide ions: Binding affinities, protein unfolding and conformational changes. *Biochemie* 2016; 123:117–129.
 45. Kumar A, Ali M, Ningthoujam RS, Gaikwad P, Kumar M, Nath BB, Pandey BN. The interaction of actinide and lanthanide ions with hemoglobin and its relevance to human and environmental toxicology. *Journal of Hazardous Materials* 2016; 307: 281–293.
 46. Muller D, Houpert P, Cambar J, Hengen-Napoli MH. Role of the sodium- dependent phosphate co-transporters and of the phosphate complexes of uranyl in the cytotoxicity of uranium in LLC-PK1 cells. *Toxicology Applied Pharmacology* 2006; 214:166-177.
 47. Liu F, Du KJ, Fang Z, You Y, Wen GB, Lin YW. Chemical and biological insights into uranium-induced apoptosis of rat hepatic cell line. *Radiation and Environmental Biophysics* 2015; 54:207-216. DOI: 10.1007/s00411-015-0588-3.
 48. Camalier CE, Young MR, Bobe G, Perella CM, Colburn NH, Beck Jr. GR. 2010. Elevated phosphate activates N-ras and promotes cell transformation and skin tumorigenesis. *Cancer Prevention Research (Phila)* 2010; 3:359-70. doi:10.1158/1940-6207.CAPR-09-0068.
 49. Konietzka R. Gastrointestinal absorption of uranium compounds - a review. *Regulatory Toxicology and Pharmacology* 2015; 71: 125-33, 2015. DOI: 10.1016/j.yrtph.2014.08.012.
 50. Sela H, Karpas Z, Zoriy M, Pickhardt C, Becker JS. Biomonitoring of hair samples by laser ablation inductively coupled plasma mass spectrometry (LA-ICP- MS). *International Journal Mass Spectrometry* 2006; 261:199-207.
 51. Sullivan M F, Ruemmler PS, Ryan JL, Buschbom RL. Influence of oxidizing or reducing agents on gastrointestinal absorption of U, Pu, Am, Cm and Pm by rats. *Health Physics* 1986; 50:223-232.
 52. Hassoun R. 2012. A statistical evaluation of the contribution of mineral and tap water to the dietary intake of As, B, Cu, Li, Mo, Ni, Pb, U and Zn by humans. Braunschweig, Germany; PhD thesis, Technical University 2020.
 53. Eastwood M. Principles of human nutrition. 2nd edition Weinheim: Wiley-Blackwell; 2003.
 54. Schnug E, Hassoun R, Holzhausen K, Jacobs F, Haneklaus SH. Contribution of mineral and tap water to the dietary intake of

- As, B, Ca, Ce, Cu, F, La, Li, Mo, Ni, P, Pb, Sr, U and Zn by humans. In: Grumezescu A, Butu A, editors. *Bottled and Packaged Water*. 2019; 1st edition, Volume 4: *The Science of Beverages*. p. 241-274. DOI:10.1016/B978-0-12-815272-0.00010-6.
55. Berdanier CD. *Handbook of Nutrition and Food*, Boca Raton: CRC Press, 2002.
56. Vanevijvere S, Jaacks LM, Monteiro CA, Moubarac J-C, Girling-Butcher M, Lee AC, Pan A, Bentham J, Sinburn B. Global trends in ultraprocessed food and drink product sales and their association with adult body mass index trajectories. *Obesity Reviews* 2019; 20:10-19.